

METHYL MERCURY TOXICOKINETICS IN CHANNEL CATFISH (ICTALURUS PUNCTATUS) AND LARGEMOUTH BASS (MICROPTERUS SALMOIDES) AFTER INTRAVASCULAR ADMINISTRATION

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Abstract—We compared the differences in the distribution and elimination of CH₃Hg after intraarterial injection and serial blood removal in catfish and bass of similar body size under consistent water quality conditions. The blood and plasma concentration—time profiles of individual fish were analyzed using a three-compartment, clearance–volume model. The plasma protein binding of CH₃Hg was determined by ultrafiltration (30,000 mol. wt. cutoff) and the binding affinity (ρ) of CH₃Hg for red blood cells (RBCs) was also calculated. Toxicokinetic analysis of the plasma concentration—time profiles provided the following values: apparent volume of distribution at steady state (V_{ss}) = 30 ± 14 ml/g (catfish), 6.2 ± 2 ml/g (bass); total body clearance (Cl_b) = 0.026 ± 0.011 ml/h/g (catfish), 0.0057 ± 0.001 ml/h/g (bass). The values of V_{ss} and Cl_b estimated from the blood concentration—time profiles in catfish and bass were fivefold lower. The elimination half-life from blood and plasma was between 814 and 1670 h and was not statistically different between species or reference fluid. The AUC_{0-xx} for blood was over three times higher than plasma, due to the binding of CH₃Hg to RBCs. The unbound fraction of CH₃Hg in bass plasma was 14-fold lower (0.25 vs. 3.64%) and the ρ for RBCs was 20 times greater than catfish (5,974 vs. 289). The decreased binding to plasma and RBCs in catfish is consistent with the increased extravascular distribution and clearance capacity of CH₃Hg in catfish because a larger fraction of the CH₃Hg in blood is available to distribute outside the vascular system.

Keywords—Methyl mercury Toxicokinetics Catfish Bass Plasma binding

INTRODUCTION

Mercury input to aquatic environments from industrial sources will increase into the next century [1,2]. As a result, the concentration of Hg in many aquatic organisms will probably continue to increase. The primary chemical form of Hg found in aquatic organisms is the highly toxic methyl mercury (CH₃Hg) [3]. This represents both an ecological and human health concern because consumption of fin and shellfish is the primary source of Hg to humans and many piscivorous wildlife [4–8]. A recent survey of state fish consumption advisories found that Hg was the most commonly cited pollutant, accounting for 60% of all advisories [9].

The accumulation of CH₃Hg in fish has been studied after water and oral exposures [10–12]. Methyl mercury is rapidly absorbed from water by rainbow trout (*Oncorhynchus mykiss*) with an uptake efficiency of 8% [13]. The assimilation of CH₃Hg from food also proceeds with high efficiency and has been reported to be 70 to 90% in goldfish (*Carassius auratus*) [14] and greater than 73% in rainbow trout [15]. The elimination of CH₃Hg from fish is slow with biological half-lives ranging from 16 to 1,000 days [12]. The long elimination half-life for CH₃Hg has been attributed primarily to its accumulation and persistence in skeletal muscle [15]. A characteristic of CH₃Hg distribution that can influence its elimination is its binding to red blood cells (RBCs). The RBC-to-plasma ratio of CH₃Hg measured after in vitro incubation with rainbow trout blood is 9 [16] and is similar to values reported for

humans [17], although substantial interspecies differences exist among other mammalian species [18].

The influence of environmental and biological variables (e.g., body size, age, acclimation temperature, water quality, and pH) on the accumulation of Hg in fish has been reported in the literature. In general, larger, older fish in low alkaline, low pH lakes tend to accumulate the highest concentrations of Hg [4,14,19,20]. Mercury body burden can vary widely among fish species within a lake [21], and this difference is often attributed to the trophic level of the species, with top predators such as largemouth bass (*Micropterus salmoides*) accumulating more Hg than forage species [21,22].

A less-studied aspect of mercury accumulation in fish is the importance of intrinsic differences among fish species, unrelated to trophic level. These differences may also influence the accumulation or clearance of Hg. The present study compares the differences in the distribution and elimination of CH₃Hg after intravascular administration to channel catfish (*Ictalurus punctatus*) and largemouth bass of similar body size under consistent water quality conditions. The plasma protein and red blood cell binding and the total lipid content of the catfish and bass are also reported to ascertain if they can be used to explain differences in the toxicokinetics between catfish and bass.

MATERIALS AND METHODS

Fish and water quality

Channel catfish (body weight \pm SD: 755 \pm 80 g) of mixed sex were obtained from Orangeburg Aquaculture (Cordova, SC, USA). Largemouth bass (696 \pm 130 g) were collected by hook and line from Par Pond, a reservoir located on the U.S.

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Department of Energy's Savannah River Site, South Carolina, USA. Immediately upon arrival at the laboratory, all fish received a 2-h treatment in a 0.25-mg/L solution of malachite green (Sigma Chemical, St. Louis, MO, USA). Catfish were held indoors in 400- and 600-L, recirculating water, fiberglass aquaria (LS 700, LS 900, Frigid Units, Toledo, OH, USA) containing reconstituted hard water [23] and 1% (w/v) NaCl. Bass were initially held outdoors in 2,800-L, polyethylenelined, circular tanks for a minimum of 2 weeks before transfer to the indoor aquaria. After transfer indoors, all fish were held at 21°C for a minimum of 2 weeks prior to use in experiments. The loading density of fish in all aquaria was maintained below 5 g/L. Half of the aquarium water was replaced biweekly. Chemical characteristics of the freshly prepared water were: total alkalinity 110 to 120 mg/L (as CaCO₃), hardness 160 to 180 mg/L (as CaCO₃), and pH 7.9. Temperature and pH were monitored daily and ranged from 20 to 22°C and 7.7 to 7.9, respectively. Ammonia levels were regularly monitored to ensure that the concentration remained below 0.5 mg/L. Catfish were fed a maintenance ration of approximately 2% of their body mass three times per week with soft moist pelleted feed (Rangen Inc., Buhl, ID, USA). Bass were fed live minnows two to three times per week.

Surgical procedures and blood removal

Each fish was fitted with a dorsal aortic cannula using methods described previously [24,25]. Catfish were anesthetized with 150 ml/L MS-222, largemouth bass with 100 mg/L MS-222. The cannula material was 28-G Teflon tubing (Zeus Inc., Raritan, NJ, USA). An 18-G intravenous catheter (Angiocath®, Becton Dickinson, Sandy, UT, USA) was used to guide the cannula into the dorsal aorta. The cannulated fish were held in 100-L polyethylene cages that were perforated to allow water exchange. The cages containing cannulated fish were placed in round, recirculating water, 1,000-L plastic aquaria filled with reconstituted hard water and 1‰ (w/v) NaCl. The end of the cannula was attached to a 1-ml syringe that floated above the fish inside the cage. All fish were allowed to recover from surgery for a minimum of 24 h before dosing.

Immediately prior to injection with CH₃Hg, a blood sample was removed for determination of hematocrit (hct) and background CH₃Hg concentrations in blood and plasma. Hematocrit was determined using heparinized microhematocrit tubes. Hematocrit was determined for all blood samples removed after CH₃Hg injection, beginning with the 4-h sample.

Catfish and bass received a 0.47-mg/kg intraarterial (IA) bolus injection of CH₃HgCl₂ (purity >99%, ICN Biomedicals, Inc., Costa Mesa, CA, USA) dissolved in 0.9% NaCl and 5 mM sodium carbonate. This dose was chosen after preliminary studies revealed that a higher dose of 1 mg/kg produced necrotic lesions around the trunk kidney and posterior skeletal muscle regions of the fish. These lesions typically appeared between 100 and 400 h after IA injection. The lower dose of 0.47 mg/kg produced no visible signs of toxicity in catfish and bass. After CH₃Hg injection, serial blood samples were removed via the cannula at 0.167, 0.33, 0.5, 0.75, 1, 1.5, 2, 4, 6, 9, 12, 24, 36 (catfish only), 50, 75, 100, 125, 200, and 300 h, and replaced with an equal volume of a modified teleost saline [26]. The volume of blood removed for each sample varied between 0.1 and 0.25 ml. The cannula was removed while the fish was under anesthesia after the 300-h sample collection. Later blood samples were obtained from the dorsal aorta of anesthetized catfish and bass using a 25-G needle

attached to a 1-ml syringe. The body weight of the fish was also recorded at this time. After 300 h, the sampling interval for blood removal was increased to approximately 250 h because CH3Hg concentrations in blood and plasma were declining slowly. Sampling continued for 3,000 h (125 d) for catfish and 1,500 h (62 d) for bass. The catfish and bass were not fed during the initial 300-h sampling period. Thereafter, catfish were fed approximately 2% of their body mass 3 d per week with the pelleted feed. This ration level was sufficient to maintain the catfish within 90% of their initial body weight. We attempted to feed the bass three times per week with live minnows. However, some individuals would not consistently feed and the final body weight of the nine bass used in this study was $82 \pm 8\%$ (mean \pm SD) of the initial body weight. Blood samples were not removed on days that fish were fed. The cumulative volume of blood removed via the cannula was less than 10% of the estimated blood volume of the fish (assumed to be 4% of body weight). After resumption of feeding, the removal of blood was limited to 0.40 to 0.75 ml per 250-h sample interval, which was less than 2% of the estimated blood volume.

Analytical procedures

Both whole blood and plasma (obtained by centrifuging blood at 2,000 g for 5 min) were assayed for CH₃Hg using a Perkin-Elmer (model 5100ZL) graphite furnace atomic absorption spectrometry (GFAAS) and a modification of methods described by Filippelli [27]. Briefly, aliquots of fresh blood or plasma (maximum sample volumes of 0.75 ml) were mixed with an equal volume of a 1% NaCl (w/v) and 2 N HCl solution, and immediately frozen. Later, the samples were thawed, and the CH3Hg was removed by quadruplicate extractions with 300 µl of benzene. The benzene extracts were pooled and then extracted with 500 µl of 2.5 mM Na₂S₂O₄ (>99.999% pure, Aldrich Chemical, Milwaukee, WI, USA). The Na₂S₂O₄ layer was removed and the CH₃Hg concentration determined by GFAAS. A standard curve of CH₃Hg in 2.5 mM Na₂S₂O₄ was prepared each day of analysis. A set of three blood or plasma standards, spiked with a known amount of CH₃Hg to encompass the expected concentration range of unknown samples, was prepared and frozen as described previously. These standards were thawed and assayed simultaneously with experimental samples. The recovery of CH₃Hg from spiked samples was (mean \pm SD) blood, 93 \pm 7%, n =15; plasma; 95 \pm 13%, n = 62. The calculated CH₃Hg concentrations were not corrected for recovery because of the consistently high recovery of CH₃Hg from spiked standards.

The blood sampling protocol was chosen due to limitations on the volume of blood that can be removed from the fish and as a compromise between the needs for frequent sampling to fully characterize the blood/plasma concentration-time profiles and high precision in the estimation of blood and plasma CH₃Hg concentrations. The final protocol was based, in part, on results of an experiment performed to determine if any advantage in analytical precision could be obtained by removing replicate blood samples at fewer sample times. A single catfish (body wt. = 950 g) was administered a 0.47-mg/kg IA dose of CH₃Hg as described previously. At 100 h postinjection, five sequential 1-ml blood samples were rapidly removed using separate, 1-ml syringes. After removal of each 1-ml sample, the cannula was rinsed with heparinized saline, then blood was redrawn into the cannula and the first 50 µl discarded before removal of the next sample. Next, plasma was obtained from each blood sample and three equal volume aliquots were separated and assayed for CH₃Hg as described previously. A nested ANOVA was performed to determine the variability associated with sample removal, extraction of CH₃Hg from plasma, and GFAAS analytical steps. Minimal (<10%) variation among plasma samples indicated there was no advantage to taking more than one sample at each sampling time.

The hct of blood samples removed via the cannula during the initial 300-h sampling period were typically less than the hct of later blood samples removed after anesthetizing the fish. This observation is consistent with previous reports indicating that the blood sampling protocol influences the hct [28]. Because the majority of CH₃Hg in blood was associated with the RBC fraction, we adjusted the blood concentrations of CH₃Hg of individual fish to the hct value determined prior to injection by the following equation:

$$C_{\text{adj}} = (C_{\text{RBC}}/\text{hct}) \cdot \text{hct}_{\text{init}} + [C_{\text{p}} \cdot (1 - \text{hct}_{\text{init}})]$$

where $C_{\rm adj}=$ adjusted blood concentration, $C_{\rm RBC}=$ the difference between the observed CH₃Hg concentrations in blood and plasma, hct = the observed hematocrit of the blood sample, hct_{init} = the hct of blood prior to injection, and $C_{\rm p}=$ the observed plasma concentration. This adjustment assumes that actual changes in hct due to blood removal during the experiment are similar between the two species.

Plasma protein and blood binding

Plasma protein binding was determined using the Centrifree micropartition system (mol. wt. cutoff 30,000; Amicon Inc., Beverly, MA, USA). Plasma samples (0.55 ml) from each species were combined with 1 μ g CH₃Hg (dissolved in 10 μ l 0.9% w/v, NaCl) and incubated at 21°C for 20 min. After incubation, a 50- μ l aliquot was removed to determine plasma CH₃Hg concentrations. The remaining plasma was pipetted into Centrifree units and centrifuged at 1,500 g for 15 min. The ultrafiltrate (100–150 μ l) and the plasma aliquot were analyzed for CH₃Hg as described previously. Calculation of the unbound fraction was made using the following equation:

% unbound = ([CH₃Hg]ultrafiltrate/[CH₃Hg]plasma) \times 100

The binding of CH_3Hg to the ultrafiltration membrane was determined to be less than 10%.

Calculation of the affinity constant of CH₃Hg for RBCs (ρ) was made using the following equation [29]:

$$\rho = [(hct - 1) + (C_b/C_p)]/(f_u \cdot hct)$$
 (1)

where $f_{\rm u}=$ the unbound plasma fraction of CH₃Hg, and $C_{\rm b}$ and $C_{\rm p}$ are the observed blood and plasma concentrations of CH₃Hg, respectively. The ratios of the AUC_{0-∞}blood/AUC_{0-∞}plasma were used to estimate $C_{\rm b}/C_{\rm p}$.

Lipid content of catfish and bass

The nonpolar lipid content of catfish and bass used in this study was determined by Soxhlet extraction of dried fish sample [30]. The fish carcass was homogenized with an equal volume of water using a Waring blender, and a 50-g aliquot was freeze dried. Three grams (dry weight) of the freeze-dried samples was used for nonpolar lipid extraction. Nonpolar lipids were extracted for a minimum of 5 h using petroleum ether. The mass of nonpolar lipids in each sample was calculated as the difference in sample dry mass before and after extraction.

Toxicokinetic analysis

An iterative, nonlinear least-squares computer program, PCNONLIN (Statistical Consultants, Inc., Lexington, KY, USA) was used to fit the blood and plasma concentration time (C_bt, C_pt) profiles of CH_3Hg after IA injection. Preliminary analysis of the C_bt , C_pt profiles using a two- or three-compartment pharmacokinetic model indicated the three-compartment model provided a better fit to the observed data based on several criteria: Akaike information criterion, coefficients of variation, and visual inspection of the fits [31,32]. The following triexponential equation was used to fit the observed blood and plasma concentration—time profiles:

$$C_{\rm b}t, \ C_{\rm p}t = Ae^{-\pi t} + Be^{-\alpha t} + Ce^{-\beta t}$$
 (2)

The area under the curve $(AUC_{0\to\infty})$, was estimated using the following equation: $AUC_{0\to\infty} = A/\pi + B/\alpha + C/\beta$. The biological half-life $(t_{1/2,\beta})$, mean residence time (MRT), area under the moment curve (AUMC), total body clearance (Cl_b) , and apparent volume of distribution at steady state (V_{ss}) were estimated by the following equations: $t_{1/2,\beta} = 0.693/\beta$, MRT = $AUMC/AUC_{0\to\infty}$, where $AUMC = A/\pi^2 + B/\alpha^2 + C/\beta^2$, $Cl_b = dose/AUC_{0\to\infty}$, $V_{ss} = Cl_b \cdot MRT$.

The fraction of the injected dose remaining in the fish (X_f) was estimated by the following equation [33]:

$$X_{f} = \frac{(A/\pi)e^{-\pi t} + (B/\alpha)e^{-\alpha t} + (C/\beta)e^{-\beta t}}{A/\pi + B/\alpha + C/\beta}$$
(3)

A simulation comparing the fraction of the injected dose of CH₃Hg remaining in catfish and bass was performed using Equation 3 and the values of A, B, C, π , α , and β determined from a fit of the averaged blood concentration—time profiles for catfish and bass.

Student's *t* test was used to test for significant differences between parameter values after using Cochran's test to determine if the variances among classes were unequal.

RESULTS

Toxicokinetic analysis of CH₃Hg in channel catfish and largemouth bass

The elimination of CH3Hg from blood and plasma was qualitatively similar for catfish and bass. After injection, the concentrations of CH₃Hg in blood and plasma declined rapidly during the first 12 h and then decreased more slowly (Figs. 1 and 2). After 125 h, the CH₃Hg concentrations began to decline in an apparent, log-linear fashion. Immediately after injection (0.167 h), the concentration of CH₃Hg in whole blood was three- to sevenfold higher than in plasma, indicating greater accumulation of CH₃Hg by blood cells (Figs. 1 and 2). The CH3Hg concentration ratio for blood and plasma remained relatively constant throughout the experiment and indicated that equilibrium of CH₃Hg was rapidly achieved between blood cells and plasma. No detectable CH3Hg was observed in catfish blood or plasma prior to injection. In bass, CH3Hg could not be detected in plasma; however, some fish had preinjection blood concentrations of CH₃Hg between 1 and 2 ng/ml.

The parameter estimates from the nonlinear, least-squares fitting of the data are shown in Table 1. The two pharmaco-kinetic parameters that are particularly useful in understanding the distribution and elimination of CH_3Hg are the apparent volume of distribution at steady-state (V_{ss}) and total body clearance (Cl_b). The V_{ss} relates the amount of CH_3Hg in the fish to the concentration in the reference fluid (blood or plasma) when

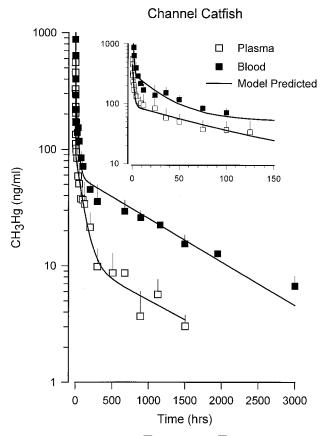


Fig. 1. The CH_3Hg plasma (\square) and blood (\blacksquare) concentration—time profiles after intraarterial injection of 0.47 mg/kg CH_3Hg in channel catfish. Symbols represent experimentally determined values (mean \pm SD, n=4-6) and the line is the least-squares fit to Equation 2. Error bars not shown fit within the data point.

an equilibrium exists between peripheral tissues and the reference fluid (i.e., steady-state conditions). The very large $V_{\rm ss}$ in catfish and bass relative to blood volume (ca. 4% body weight) [34] indicated that CH₃Hg was extensively distributing outside the vascular system (Table 1). The $Cl_{\rm b}$ represents the sum of all elimination pathways (i.e., urinary, fecal, branchial,

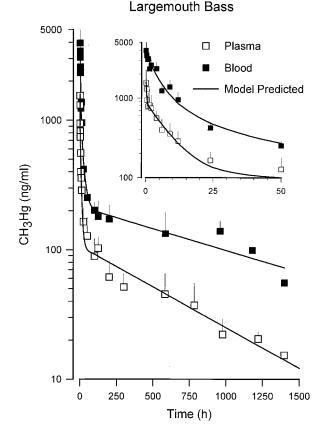


Fig. 2. The CH_3Hg plasma (\square) and blood (\blacksquare) concentration—time profiles after intraarterial injection of 0.47 mg/kg CH_3Hg in largemouth bass. Symbols represent experimentally determined values (mean \pm SD, n = 5-8) and the line is the least-squares fit to Equation 2. Error bars not shown fit within the data point.

and metabolic) of CH_3Hg from catfish and bass and can be interpreted as the volume (ml) of blood or plasma that is completely cleared of CH_3Hg per hour. Thus, the small Cl_b (Table 1) relative to cardiac output (0.06–0.7%) in catfish [28] and other teleosts [35,36] indicated only a negligible quantity of CH_3Hg is excreted relative to the flow rate of blood through

Table 1. Toxicokinetic parameters determined from the plasma and blood concentration-time profiles of channel catfish and largemouth bass

	Channel catfish ^a		Largemouth bass ^b	
Parameter	Plasma	Blood	Plasma	Blood
$\begin{array}{ c c }\hline & & & \\ & AUC_{0\to\infty} \ (ng/ml \ h) \\ & C_{max} \ (ng/ml) \\ & V_{ss} \ (ml/g) \\ & Cl_b \ (ml/h/g) \\ & t_{1/2} \ (h) \\ & MRT \ (h) \\ & f_u \ (\%)^c \\ \end{array}$	23,300 ± 7,200** 624 ± 47 30 ± 14** 0.0260 ± 0.0110* 1,670 ± 683 1,829 ± 953 3.64 ± 0.35	77,500 ± 9,100** 6,050 ± 1,950 6.2 ± 1* 0.0059 ± 0.0001** 814 ± 58 1,046 ± 77	97,600 ± 34,200 2,361 ± 382 6.2 ± 2 0.0057 ± 0.001 905 ± 189 1,054 ± 324 0.25 ± 0.02	348,600 ± 84,000 4,400 ± 450 1.6 ± 0.1 0.0014 ± 0.0004 1,008 ± 275 1,261 ± 349
$ ho^{ m d}$		289		5,974

Catfish and bass were administered a 0.47-mg/kg intraarterial dose of CH₃Hg. Parameter estimates were calculated by the PCNONLIN program after fitting the individual blood and plasma concentration–time profiles to Equation 2 to obtain estimates of A, B, C, π , α , and β . The mean \pm SD is presented for the following samples sizes: catfish, n=6 plasma, n=4 blood; bass, n=8 plasma, n=5 blood. Asterisks indicate values significantly different from bass: *p<0.05; **p<0.01.

^a Mean body weight (\pm SD) = 755 \pm 80 g.

^b Mean body weight (\pm SD) = 696 \pm 130 g.

 $^{^{}c}f_{u}$ is the percentage of CH $_{3}$ Hg that is unbound to plasma proteins and was determined by ultrafiltration of plasma through a 30,000-mol. wt. cutoff membrane.

^d ρ is the affinity constant for CH₃Hg to RBCs as calculated with Equation 1.

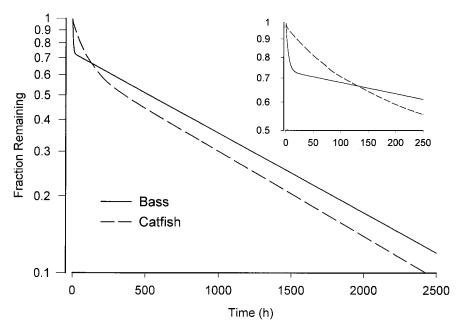


Fig. 3. Simulation depicting the fraction of an intraarterial injected dose of CH_3Hg remaining in the fish. The fraction remaining was calculated using Equation 3 and the values of A, B, C, π , α , and β determined from the fit of the averaged blood concentration—time profiles shown in Figures 1 and 2.

eliminating organs. It is the combination of a large $V_{\rm ss}$ and small $Cl_{\rm b}$ that causes the long biological half-life of CH₃Hg (>34 d; Table 1), whether estimated from the blood or plasma concentration–time profiles.

The effects of binding to blood cells by CH₃Hg was evident in several parameter estimates. The AUC for blood was three to four times higher than plasma in both catfish and bass, and the estimates of the $V_{\rm ss}$ and $Cl_{\rm b}$ were approximately fourfold lower from blood (Table 1). The $t_{\rm 12,\beta}$ and MRT were similar in catfish and bass although the plasma $t_{\rm 12,\beta}$ estimated for catfish was higher, but not significantly so (Table 1).

The most statistically significant differences in parameter estimates between catfish and bass were the larger $V_{\rm ss}$ and higher $Cl_{\rm b}$ observed in catfish (Table 1). The plasma referenced estimates of the $V_{\rm ss}$ and $Cl_{\rm b}$ were 4.8 and 4.5 times larger in catfish compared to bass, respectively (Table 1). These differences indicated that in catfish, ${\rm CH_3Hg}$ was concentrated to a greater extent in peripheral tissues, and that catfish have a greater capacity for elimination of ${\rm CH_3Hg}$ compared to bass. A simulation of the fraction of an IA dose of ${\rm CH_3Hg}$ remaining in the fish is shown in Figure 3. The simulation predicts that initially a greater fraction of the injected dose is retained by catfish, but later (>250 h), a larger fraction is retained by bass.

The mean lipid contents of catfish and bass used in this study were 3.6 ± 0.8 and $1.4 \pm 0.7\%$, respectively. Cursory examination of these data indicated no obvious correlation between lipid content and differences in the toxicokinetic parameters.

Plasma and blood binding

There were substantial differences between catfish and bass in the plasma protein binding and RBC binding affinity of CH₃Hg. The unbound fraction in plasma was more than 14-fold greater in catfish compared to bass (Table 1). The affinity constant of CH₃Hg for bass RBCs (ρ), was over 20 times greater than catfish (Table 1). The pre-injection hcts were similar among the two species and were 23 \pm 3 and 26 \pm 3%

(mean \pm SD; n=4 catfish, n=5 bass; Table 1) for bass and catfish, respectively. There was little change in the hct during the initial 100 h after injection; however, the hct was approximately 60% of the initial hct for both species by 300 h (data not shown).

DISCUSSION

The results of this study are consistent with previous reports of the slow elimination and long biological half-life of CH₃Hg in fish. Despite the similarities in biological half-life for both species, significant species differences in the V_{ss} and Cl_b estimates were observed (Table 1). The differences in V_{ss} and Cl_b appeared to be explained by the lower plasma protein and RBC binding of CH₃Hg in catfish. For chemicals that are slowly excreted like CH₃Hg, changes in blood and plasma binding will directly affect the clearance with decreased binding increasing the clearance [37]. Similarly, changes in plasma binding can also directly affect the volume of distribution if there is little change in tissue binding [38]. Although catfish used in this study had a higher body lipid content than bass, little correlation between lipid content and CH3Hg accumulation has been demonstrated [3]. Therefore, the greater V_{ss} in catfish is probably due to decreased plasma binding and/or increased lean tissue binding. The similarity in the biological half-life of CH₃Hg despite significant species differences in Cl_b is due to the larger distributional space in catfish from which CH₃Hg must be cleared.

Mercury has an extremely high affinity for sulfhydryl groups [39] including those of low molecular weight such as glutathione [40,41], and the conventional pharmacological concept of an unbound or free fraction probably does not apply to CH₃Hg. However, CH₃Hg rapidly exchanges between glutathione and hemoglobin in human erythrocytes [40]. Also, the average lifetime of the CH₃Hg–glutathione complex is less than 0.01 s [40,41], suggesting that the pharmacokinetic behavior of CH₃Hg bound to small molecular weight ligands may be analogous to that of an unbound molecule. In our

binding studies, the ultrafiltrate probably represents CH₃Hg bound to small molecular weight ligands that were able to pass through the membrane (30,000 mol. wt. cutoff). A similar result for CH₃Hg has been reported for human plasma, where the ultrafiltrate fraction of CH₃Hg was 0.8 and 8% after filtering plasma through membranes with mol. wt. cutoffs of 1,000 and 300,000 MU [42]. Also, zone electrophoresis of human plasma showed that only 20% of the protein-bound CH₃Hg was associated with albumin and over 45% associated with unidentified proteins of molecular weights greater than albumin [42].

The large difference in blood binding of CH_3Hg between bass and catfish (ρ ; Table 1) is similar to the greater binding of CH_3Hg to rat blood cells compared with other mammalian species [18]. The difference in blood cell binding affinity among mammalian species is not due to greater binding to hemoglobin [18] but to some unknown mechanism. Our results indicated that bass were closer to rats in their blood cell affinity for CH_3Hg and catfish are more similar to humans and other mammalian species.

The primary elimination routes of CH₃Hg from rodents and man are biotransformation to inorganic Hg and secretion of CH3Hg into bile, bound to various nonprotein sulfhydryl compounds [43-45]. Of these two excretory routes, conversion of CH₃Hg to inorganic Hg is considered more important [44]. These elimination pathways also appear to occur in fish, as a significant portion of 203Hg is recovered in the feces after CH₃²⁰³Hg administration in trout [15], and the isotope ratio of ²⁰³Hg to ¹⁴C increases over a 6-week period in the kidney and liver of trout coadministered ¹⁴CH₃Hg and CH₃²⁰³Hg [46]. The importance of CH₃Hg conversion to inorganic Hg in fish is unclear as most of the Hg in edible fish tissue (>98%) is CH₃Hg [3]. However, after IA administration of inorganic Hg to channel catfish, most of the Hg eventually becomes concentrated in the liver with only trace quantities accumulating in skeletal muscle [47]. Furthermore, the biological half-life of inorganic Hg in channel catfish exceeds 700 d [47], considerably longer than CH₃Hg (34 d, Table 1). These observations indicate that inorganic Hg is much more persistent in fish and the accumulation of inorganic Hg in the liver after CH₃Hg exposure may provide an estimate of CH₃Hg conversion to inorganic Hg.

An additional elimination pathway for CH₃Hg may be branchial. Methyl mercury is lipophilic [48] and rapidly absorbed from water across the gills [13,49]. Assuming the ultrafiltrate or unbound fraction of CH₃Hg in plasma is available to diffuse across the gills, then comparison of the unbound Cl_b (Cl_b/f_u) values relative to gill blood flow (assumed to equal cardiac output) can provide an estimation of the importance of branchial elimination. The cardiac output in channel catfish at 21°C is 2.4 ml/h/g [28]. The cardiac output in largemouth bass has not been reported, but values for other teleosts range from 1.1 to 3.75 ml/h/g [35,36]. The unbound Cl_h for catfish and bass was 0.71 and 2.28 ml/h/g respectively, which is approximately 30% of cardiac output in catfish and probably a higher percentage for bass. This high, unbound Cl_b relative to cardiac output does not rule out biliary excretion and metabolism as important elimination pathways, but it does suggest that branchial excretion of CH3Hg may be an additional excretory route for CH₃Hg in fish.

The effects of the differences in $V_{\rm ss}$ and $Cl_{\rm b}$ on the pattern of CH₃Hg elimination is illustrated in Figure 3. This simulation predicts that a greater fraction of an IA dose is initially excreted

by bass when the blood concentration is higher than in catfish. However, by the end of the distributional phase (150–250 h), a larger portion of the dose has been excreted by the catfish (Fig. 3). The slopes of the terminal portion of the curves are similar due to the similarity in biological half-lives (Table 1).

Although the model predicts that a greater fraction of an IA-administered dose will be retained by bass (Fig. 3), the body burdens of CH₃Hg in naturally exposed bass and catfish would depend on the exposure conditions and assimilation efficiency from food and water, assuming other factors are equal (i.e., age, body size). Our analysis indicates that under identical exposure conditions, the body burden of CH₃Hg in catfish and bass of similar age and size should be equivalent due to similarities in biological half-life, but that blood and plasma concentrations of CH₃Hg will be higher in bass due to the greater binding to RBCs and plasma proteins.

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REFERENCES

- Watson, W.D. 1979. Economic considerations in controlling mercury pollution. In J.O. Nriagu, ed., *The Biogeochemistry of Mercury in the Environment*. Elsevier-North Holland, Amsterdam, The Netherlands, pp. 42–77.
- Nriagu, J.O. and J.M. Pacyna. 1988. Quantitative assessment of worldwide contamination of air, water and soils by trace metals. *Nature* 333:134–139.
- Bloom, N. 1992. On the chemical form of mercury in edible fish and marine invertebrate tissue. *Can. J. Fish. Aquat. Sci.* 49:1010– 1140.
- 4. **Weiner, J.G.** 1987. Metal contamination of fish in low-pH lakes and potential implications for piscivorous wildlife. *Trans. N. Am. Wildl. Nat. Resour. Conf.* **52**:645–657.
- Tollefson, L. and F. Cordle. 1986. Methylmercury in fish: A review of residual levels, fish consumption and regulatory action in the United States. *Environ. Health Perspect.* 68:203–208.
- Fitzgerald, W.F. and T.W. Clarkson. 1991. Mercury and monomethylmercury: Present and future concerns. *Environ. Health Perspect.* 96:159–166.
- Sarmani, S.B., A.Z. Kiprawi and R.B. Ismail. 1994. Mercury determination in hair of malaysian fisherman by neutron activation analysis. *Biol. Trace Element Res.* 40:435–441.
- Fleming, L.E., S. Watkins, R. Kaderman, B. Levin, D.R. Ayyar, M. Bizzlo, D. Stephens and J.A. Bean. 1995. Mercury exposure in humans through food consumption from the everglades of Florida. Water Air Soil Pollut. 80:41–51.
- Cunningham, A., S.L. Smith, J.P. Tippett and A. Greene. 1994.
 A national fish consumption advisory data base: A step toward consistency. Fisheries 19:14–23.
- McKim, J.M., G.F. Olson, G.W. Holcombe and E.P. Hunt. 1978. Long-term effects of methylmercuric chloride on three generations of brook trout (*Salvelinus fontinalis*): Toxicity, accumulation, distribution, and elimination. *J. Fish Res. Board Can.* 33:2726–2739.
- 11. **Pentreath, R.J.** 1976. The accumulation of mercury from food by the plaice, *Pleuronectes platessa* L. *J. Exp. Mar. Biol. Ecol.* **25**:51–65.
- Huckabee, J.W., J.W. Elwood and S.G. Hildebrand. 1979. Accumulation of mercury in freshwater biota. In J.O. Nriagu, ed., *The Biogeochemistry of Mercury in the Environment*. Elsevier/ North-Holland, New York, NY, USA, pp. 277–302.
- Boddington, M.L., B.A. Mackenzie and A.S.W. DeFrietas. 1979. A respirometer to measure the uptake efficiency of water-borne contaminants in fish. *Ecotoxicol. Environ. Saf.* 3:383–393.
- Sharpe, M.A., A.S.W. DeFrietas and A.E. McKinnon. 1976.
 The effect of body size on methylmercury clearance by goldfish (*Carassius auratus*). Environ. Biol. Fish. 2:177–183.

- Giblin, F.J. and E.J. Massaro. 1973. Pharmacodynamics of methyl mercury in the rainbow trout (*Salmo gairdneri*): Tissue uptake, distribution and excretion. *Toxicol. Appl. Pharmacol.* 24: 81–91.
- 16. Olson, K.R., H.L. Bergman and P.O. Fromm. 1973. Uptake of methyl mercuric chloride and mercuric chloride by trout: A study of uptake pathways into the whole animal and uptake by erythrocytes in vitro. J. Fish. Res. Board Can. 30:1293–1299.
- Suzuki, T., T. Miyama and H. Katsunama. 1970. Mercury contents in the red blood cells, plasma, urine, and hair from workers exposed to mercury vapor. *Ind. Health* 8:39–47.
- White, J.F. and A. Rothstein. 1973. The interaction of methyl mercury with erythrocytes. *Toxicol. Appl. Pharmacol.* 26:370– 384
- Cope, W.G., J.G. Wiener and R.G. Rada. 1990. Mercury accumulation in yellow perch in Wisconsin seepage lakes: Relation to lake characteristics. *Environ. Toxicol. Chem.* 9:931–940.
- Ponce, R.A. and N.S. Bloom. 1991. Effect of pH on the bioaccumulation of low level, dissolved methylmercury by rainbow trout (Onchorhynchus mykiss). Water Air Soil Pollut. 56:631– 640
- Lange, T.R., H.E. Royals and L.L. Connor. 1994. Mercury accumulation in largemouth bass (*Micropterus salmoides*) in a Florida lake. Arch. Environ. Contam. Toxicol. 27:466–471.
- Allard, M. and P.M. Stokes. 1989. Mercury in crayfish species from thirteen Ontario lakes in relation to water chemistry and smallmouth bass (*Micropterus dololmieui*) mercury. *Can. J. Fish. Aquat. Sci.* 46:1040–1046.
- 23. U.S. Environmental Protection Agency. 1978. Quality assurance guidelines for biological testing. EPA-600/4-78-043. Environmental Monitoring Series. Environmental Monitoring and Support Laboratory, Cincinnati, OH.
- Schultz, I.R. and W.L. Hayton. 1993. The toxicokinetics of trifluralin in rainbow trout. Aquat. Toxicol. 26:287–306.
- Schultz, I.R., W.L. Hayton and B.H. Kemmenoe. 1995. Disposition and toxicokinetics of diquat in channel catfish. *Aquat. Toxicol.* 33:297–310.
- Houston, A.H. 1990. Blood and circulation. In C.B. Schreck and P.B. Moyle, eds., *Methods for Fish Biology*. American Fisheries Society, Bethesda, MD, USA, pp. 273–322.
- 27. Filippelli, M. 1987. Determination of trace amounts of organic and inorganic mercury in biological materials by graphite furnace atomic absorption spectrometry and organic mercury speciation by gas chromatography. *Anal. Chem.* 59:116–119.
- McKim, J.M., J.W. Nichols, G.J. Lien and S.L. Bertelsen. 1994. Respiratory–cardiovascular physiology and chloroethane gill flux in the channel catfish *Ictalurus punctatus*. J. Fish Biol. 44:527–547.
- Rowland, M. and T.N. Tozer. 1988. Clinical Pharmacokinetics: Concepts and Applications. Lea & Febiger, Philadelphia, PA, USA, pp. 476–478.
- Williams, S. 1984. Official Methods of Analysis of the Association of Official Analytical Chemists, 14th ed. Association Official Analytical Chemists, Arlington, VA, USA.
- Boxenbaum, H.G., S. Riegelman and R.M. Elashoff. 1974. Statistical estimation in pharmacokinetics. J. Pharmacokinet. Biopharm. 2:123–149.
- 32. Yamaoka, K., N. Terumichi and T. Uno. 1978. Application of Akaike's information criterion (aic) in the evaluation of linear

- pharmacokinetic equations. J. Pharmacokinet. Biopharm. 6:165–175
- Chiou, W.L. 1972. A simple equation to estimate the fraction of drug remaining in the body after an intravenous injection. J. Pharm. Pharmacol. 34:342–344.
- 34. Olson, K.R. 1992. Blood and extracellular fluid volume regulation: Role of the renin–angiotensin system, kallikrein–kinin system, and atrial natriuretic peptides. In W.S. Hoar, D.J. Randall and A.P. Farrell, eds., *Fish Physiology*, Vol. 12B—The Cardiovascular System. Academic, San Diego, CA, USA, pp. 1–135.
- Barron, M.G., B.D. Tarr and W.L. Hayton. 1987. Temperature-dependence of cardiac output and regional blood flow in rainbow trout, Salmo gairdneri Richardson. J. Fish Biol. 31:735–744.
- Kolok, A.S., R.M. Spooner and A.P. Farrell. 1993. The effect of exercise on the cardiac output and blood flow distribution of the largescale sucker *Catostomus macrocheilus*. J. Exp. Biol. 183: 301–321
- Wilkinson, G.R. 1987. Clearance approaches in pharmacology. Pharmacol. Rev. 39:1–47.
- Gibaldi, M. and P.J. McNamara. 1978. Apparent volumes of distribution and drug binding to plasma proteins and tissues. *Eur. J. Clin. Pharmacol.* 13:373–378.
- Hughes, W.L. 1957. A physicochemical rationale for the biological activity of mercury and its compounds. *Ann. N.Y. Acad. Sci.* 65:454–460.
- Rabenstein, D.L., A.A. Isab and R.S. Reid. 1982. A proton nuclear magnetic resonance study of the binding of methylmercury in human erythrocytes. *Biochim. Biophys. Acta* 696:53–64.
- Rabenstein, D.L., A.P. Arnold and R.D. Guy. 1986. ¹H-NMR study of the removal of methylmercury from intact erythrocytes by sulfhydryl compounds. *J. Inorg. Biochem.* 28:279–287.
- Greener, Y. and J.A. Kochen. 1983. In vitro studies on methyl mercury distribution in human blood. *Teratology* 28:375–387.
- Stein, A.F., Z. Gregus and C.D. Klaassen. 1988. Species variations in biliary excretion of glutathione-related thiols and methylmercury. *Toxicol. Appl. Pharmacol.* 93:351–359.
- 44. Farris, F.F., R.L. Dedrick, P.V. Allen and J.C. Smith. 1994. Physiological model for the pharmacokinetics of methyl mercury in the growing rat. *Toxicol. Appl. Pharmacol.* 119:74–90.
- Smith, J.C., P.V. Allen, M.D. Turner, B. Most, H.L. Fisher and L.L. Hall. 1994. The kinetics of intravenously administered methyl mercury in man. *Toxicol. Appl. Pharmacol.* 128:251–256.
- 46. Olson, K.R., K.S. Squibb and R.J. Cousins. 1978. Tissue uptake, subcellular distribution, and metabolism of ¹⁴CH₃HgCl and CH₃²⁰³HgCl by rainbow trout (*Salmo gairdneri*). J. Fish. Res. Board Can. 35:381–390.
- Schultz, I.R., E.L. Peters and M.C. Newman. 1996. Toxicokinetics and disposition of inorganic mercury and cadmium in channel catfish after intravascular administration. *Toxicol. Appl. Pharmacol.* 140:39–50.
- Bienvenue, E., A. Boudou, J.P. Desmazes, C. Gavach, D. Georgescauld, J. Sandeaux, R. Sandeaux and P. Seta. 1984. Transport of mercury compounds across bimolecular lipid membranes: Effect of lipid composition, pH, and chloride concentration. *Chem. Biol.-Interact.* 48:91–101.
- Phillips, G.R. and D.R. Buhler. 1978. The relative contributions of methylmercury from food or water to rainbow trout (*Salmo gairdneri*) in a controlled laboratory environment. *Trans. Am. Fish. Soc.* 107:853–861.